

1 **Appendix A**

2 **Model parameterization (input data) details:**

3 a) **Digital elevation model** – United States Geological Survey National Elevation Dataset

4 b) **Stream channel network** - USGS National Hydrography Dataset (NHD)

5 c) **Landcover** - USGS 2001 National Land Cover Dataset

6 d) **Soil types** – Natural Resource Conservation Service State Soil Geographic Database

7 e) **Weather** – National Oceanic and Atmospheric Administration National Climatic Data Center
8 and Michigan State University’s Michigan Climatological Resources Program; daily precipitation,
9 minimum and maximum temperature, windspeed, relative humidity, and solar radiation for 150
10 local weather stations.

11 f) **Point source discharges** - EPA Permit Compliance System (PCS) database, supplemented
12 when necessary with averages from more complete local discharger datasets; average daily
13 water flow, total suspended solids (sediment), organic phosphorus, organic nitrogen, nitrate,
14 ammonia, nitrite, soluble reactive phosphorus (mineral phosphorus), chemical/biological
15 oxygen demand, and dissolved oxygen data.

16 g) **Impoundment characteristics** - Allan and Hinz (2004); USGS National Hydrography Dataset;
17 surface area and volume measurements for typical and flood stage conditions were collected.
18 Impoundments located on stream channels, including natural lakes and artificial reservoirs,
19 were included in the SWAT models. Impoundments with surface areas > 50 ha were
20 incorporated into SWAT as reservoirs, and impoundments of 10 - 50 ha were incorporated as
21 ponds. The impoundment outflows were estimated based on the targeted release option
22 which allows some extra filling of impoundments during flood events, followed by return back
23 to volume at principal spillway depth within a few days. When multiple ponds or reservoirs
24 were located in a single subwatershed, the surface area and volume were summed since SWAT
25 only accepts one pond and one reservoir per subwatershed. Water quality parameters to
26 depict P and N settling in impoundments were adjusted on the basis of a study of 27
27 Midwestern reservoirs (Walker and Kuhner, 1978) that reported median settling rates of 14.0
28 m/y for P and 4.7 m/y for N. All other reservoir and pond input parameters were set at default
29 values.

30 h) **Atmospheric N deposition** - National Atmospheric Deposition Program (NADP); unique N
31 concentration used for each watershed.

32 i) **Land management practices** - USDA National Agricultural Statistics Service (NASS), various
33 local experts, and Ruddy et al. (2006); generalized land management schedules (operations
34 include fertilizer application, tillage, crop planting/harvesting, and manure application) applied
35 to each watershed uniquely.

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37 Allan JD, Hinz L. An assessment of flows for rivers of the great lakes basin. Project report
38 prepared under grant from Great Lakes Protection Fund. Ann Arbor, MI. 2004.

39 Ruddy BC, Lorenz DL, Mueller DK. County-level estimates of nutrient inputs to the land surface
40 of the conterminous United States, 1982-2001. National Water Quality Assessment
41 Program, U.S. Geological Survey. 2006.

1 Walker WW, Kuhner J. An empirical analysis of factors controlling eutrophication in
2 midwestern impoundments. In: International Symposium on the Environmental Effects
3 of Hydraulic Engineering Works, University of Tennessee, Knoxville, TN. 1978.
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5 **Model calibration and validation details** (from Bosch et al., 2011)

6 Model calibration included several sequential stages for each individual watershed,
7 including hydrology sensitivity analysis, hydrology manual calibration, hydrology
8 autocalibration, sediment manual calibration, and nutrient manual calibration. First, an
9 automated sensitivity analysis (Van Griensven et al., 2006) is carried out through the ArcSWAT
10 interface with hydrologic model parameters in order to identify the parameters to be adjusted
11 during the autocalibration procedure as well as to give some insight into parameters to adjust
12 during manual hydrology calibration. The sensitivity analysis procedure uses the Latin
13 hypercube one factor at a time design, and identifies the top 15 most sensitive parameters.
14 Next, the hydrology was roughly calibrated manually by changing sensitive hydrologic
15 parameters as described in Santhi et al. (2001). Once simulated stream discharge roughly fit
16 the observed discharge for the calibration time period, a second sensitivity analysis was
17 performed. The top 15 parameters of both the pre- and post-manual calibration sensitivity
18 analysis runs were then used for the hydrology autocalibration. The hydrology autocalibration
19 employed the PARASOL calibration procedure included in the ArcSWAT interface (Van
20 Griensven and Meixner, 2007). The PARASOL method applied a shuffled complex evolution
21 optimization scheme to select the optimal parameter value set for the 15 sensitive hydrologic
22 parameters after several thousand model runs. The calibration was based on monthly
23 observed USGS daily mean stream discharge data (1998-2001) and SWAT simulated stream
24 discharge. Hydrologic model parameter values were then adjusted to reflect the optimal value
25 set chosen by the autocalibration process.

26 After model hydrology calibration was completed, manual calibration of sediments and
27 nutrients was completed, followed by model validation. SWAT sediment parameters were
28 calibrated following the procedure of Santhi et al. (2001) based on monthly observed sediment
29 loads from 1998-2001. Since no observed sediment data were available for the Huron River,
30 the optimized sediment parameter values from the adjacent Raisin watershed were used for
31 the Huron SWAT model as well. After sediment calibration was completed, TP, SRP, TN, and
32 nitrate calibration was done based on monthly observed nutrient data (Santhi et al., 2001).
33 SWAT output included mineral P and other P, so SWAT mineral P was compared to observed
34 SRP data and observed TP was compared to the sum of SWAT mineral P and other P. After
35 model nutrient parameters were optimized for all six watershed models, calibration was
36 complete and model validation was initiated. During model validation, evaluation statistics
37 were calculated for stream discharge, sediments, TP, SRP, TN, and nitrate and for each of the six
38 SWAT models (Moriassi et al. 2007), but no further model parameter changes were made.
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41 Water Assessment Tool for six watersheds of Lake Erie: Model parameterization and
42 calibration. *J. Gt. Lakes Res.* 37, 263-71.

1 Moriasi DN, Arnold JG, Van Liew MW, Bingner RL, Harmel RD, Veith TL. Model evaluation
2 guidelines for systematic quantification of accuracy in watershed simulations. Trans
3 ASABE 2007;50:885-900. Santhi C, Arnold JG, Williams JR, Dugas WA, Srinivasan R, Hauck
4 LM. Validation of the SWAT model on a large river basin with point and nonpoint
5 sources. J Amer Water Resour Assoc 2001;37:1169-88.

6 van Griensven A, Meixner T. A global and efficient multi-objective auto-calibration and
7 uncertainty estimation method for water quality catchment models. J Hydroinformatics
8 2007;9:277-91.

9 van Griensven A, Meixner T, Grunwald S, Bishop T, Diluzio A, Srinivasan R. A global sensitivity
10 analysis tool for the parameters of multi-variable catchment models. J Hydrol
11 2006;324:10-23.

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